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Variation in $\delta^{13}\text{C}$ among species and sexes in the family Restionaceae along a fine-scale hydrological gradient

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1 **Variation in $\delta^{13}\text{C}$ among species and sexes in the family Restionaceae along a**
2 **fine-scale hydrological gradient**

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ABSTRACT

Consistent, repeatable segregation of plant species along hydrological gradients is an established phenomenon that must in some way reflect a trade-off between plants' abilities to tolerate the opposing constraints of drought and waterlogging. In C_3 species tissue carbon isotope discrimination ($\delta^{13}C$) is known to vary sensitively in response to stomatal behaviour, reflecting stomatal limitation of photosynthesis during the period of active growth. However, this has not been studied at fine-spatial scale in natural communities. We tested how $\delta^{13}C$ varied between species and sexes of individuals in the family Restionaceae growing along a monitored hydrological gradient. 20 Restionaceae species were investigated using species-level phylogeny at two sites in the Cape Floristic Region, a biodiversity hotspot. A spatial overlap analysis showed the Restionaceae species segregated significantly ($p < 0.001$) at both sites. Moreover, there were significant differences in $\delta^{13}C$ values among the Restionaceae species ($p < 0.001$) and between male and female individuals of each species ($p < 0.01$). However, after accounting for phylogeny, species $\delta^{13}C$ values did not show any significant correlation with the hydrological gradient. We suggest that some other variable (e.g. plant phenology) could be responsible for masking a simple response to water availability.

Key words: Cape floristic region, carbon isotope discrimination, hydrological gradient, plant gender, Restionaceae, soil water regime trade-off, water use efficiency

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39 **INTRODUCTION**

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41 Plants are in general very sensitive to small differences in soil moisture, even
42 within a range of moisture tensions where water is still freely available to them
43 (Davies & Gowing 1999). Classical experiments by Ellenberg (1953) with meadow
44 grassland plants demonstrated that interspecific competition leads to segregation of
45 species' distributions along soil moisture gradients and Silvertown *et al.* (1999)
46 further showed that this structures meadow plant communities in the field.

47

48 Niche separation on fine-scale hydrological gradients may well be common in
49 plant communities, although the physiological basis of this segregation is still unclear.
50 In general, separation of species into distinct niches along any resource or
51 environmental axes is caused by trade-offs that force species to specialize (MacArthur
52 1972). Possible causes of hydrological niche specialization that have been suggested
53 include trade-offs between tolerance of aeration stress (caused by waterlogging) and
54 drought stress (Davies & Gowing 1999; Silvertown *et al.* 1999) and tolerance of low
55 soil nutrient availability (in wet conditions) and drought stress (e.g. Neill 1990;
56 Castelli *et al.* 2000; Araya 2005).

57

58 Whatever the precise trade-offs may be that underlie the specialization of
59 species into hydrological niches, these must involve the water economy of plants and
60 hence must be related to stomatal behaviour. Stomatal function imposes a
61 fundamental trade-off between water conservation and carbon assimilation (and hence
62 growth) because stomata must open to allow CO₂ uptake, but must close to limit water
63 loss. Thus water stress may cause stomatal limitation of photosynthesis during periods

of active growth (e.g. Henson *et al.* 1989). The ratio of CO₂ assimilated to stomatal conductance determines the Intrinsic Water Use Efficiency (WUE) of a plant. Could this fundamental trade-off arising from the two conflicting functions of stomata be the ultimate explanation for niche separation along hydrological gradients? If it is, then intrinsic WUE ought to vary between species in a systematic manner along soil moisture gradients.

Precisely how intrinsic WUE is expected to vary along soil moisture gradients depends upon how other influences upon growth such as soil-available nitrogen or oxygen supply vary along the gradient too. If no other factor is limiting to growth, intrinsic WUE ought to decrease monotonically with increasing water supply. If soil-available nitrogen or oxygen (or both) become limiting at the wet end of the gradient, WUE ought to reach a maximum near the middle of the gradient. In theory, these patterns ought to be detectable through changes in ¹³C isotope discrimination, a method often used for assessment of intrinsic WUE (Farquhar *et al.* 1982; Dawson *et al.* 2002). When stomatal resistance is high because of closure or increased photosynthetic demand, the influence of enzymatic discrimination is diminished and $\delta^{13}\text{C}$ leaf values are enriched. On the other hand, if resistance is reduced the relative influence of enzymatic discrimination increases, favouring ¹³C depletion during fixation (Van de Water *et al.* 2002).

In experiments with crop plants, changes in stomatal resistance and behaviour eventually led to changes in plant water use efficiency (e.g. Farquhar and Richards 1984; Martin & Thorstenson 1988; Ebdon *et al.* 1998). However, this relationship is not as straightforward as has often been assumed in the past (Seibt *et al.* 2008). This is

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89 because, particularly in natural vegetation, C_i/C_a (the ratio of CO_2 concentrations in
90 the leaf intercellular spaces to that in the atmosphere), which partly determines
91 isotope discrimination and in turn WUE, could be influenced by external biotic and
92 abiotic factors (Griffiths *et al.* 2000). Hence, WUE and $\delta^{13}C$ can vary independently
93 of one another, making the use of $\delta^{13}C$ as a surrogate for WUE questionable (Griffiths
94 *et al.* 2000, Seibt *et al.* 2008). This is particularly the case where $\delta^{13}C$ is used alone,
95 without other independent estimates of gas exchange or environmental conditions.

96
97 In this paper, we test for a correlation between $\delta^{13}C$ and the location of species
98 on a soil moisture gradient in the field, using two methods for controlling variation
99 that could obscure the relationship. First, our 20 study species all belong to the same
100 clade, the African Restionaceae, which are dioecious, graminoid C_3 perennials
101 (Linder *et al.* 2005). We selected the Restionaceae for study because many species
102 belonging to this family co-occur within fynbos habitats and species are found across
103 a wide range of water regimes (Linder *et al.* 1998, Hardy *et al.* 2008). By use of a
104 phylogeny for this clade that is resolved to species level (Hardy *et al.* 2008), we are
105 able to use phylogenetically independent contrasts in correlations between $\delta^{13}C$ and
106 other variables, thus eliminating the potential influence of phylogenetically-correlated
107 hidden third variables. Second, we make intraspecific comparisons between males and
108 females of Restionaceae species. Females are generally expected to have lower WUE
109 than males in dioecious species (Dudley 2006).

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111 **METHODS**

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113 **Site and species**

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115 Restionaceae species were sampled from two study sites, located at the heart
116 of the Cape Floristic Region, South Africa a global biodiversity hotspot (Myers *et al.*
117 2000). The rationale of site selection was to test our hypothesis in sites with
118 contrasting altitude, species composition and water regime. The first site was at
119 Riverlands Nature Reserve (33° 29' 14.2" S, 18° 35' 44.1" E) at an altitude of 120 m
120 a.s.l, while the second one was at New Years Peak, within Limietsberg Nature
121 Reserve (33° 41' 19.6" S, 19° 06' 03.1" E) at an altitude of 1085 m. The annual
122 rainfall at Riverlands was 375 mm, while for New Years Peak it was 1660 mm.
123 Potential evapotranspiration rates were 3.5 mm and 2.8 mm per day for Riverlands
124 and New Years Peak respectively.

125

126 In a plot of 11,442 m² at New Years Peak and 2500 m² size at Riverlands, 235
127 and 305 1 m² quadrats were surveyed for the presence/absence of Restionaceae
128 species, and when possible, for the sex of each plant. For the study species, samples
129 for $\delta^{13}\text{C}$ analysis were taken in replicates of ten along a transect placed through the
130 plot.

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132 Carbon isotope analysis

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134 From flowering culms of sampled species between 6-14 months old, a ten cm
135 long stem sample was taken ten cm above ground surface. The collected plant
136 material was then oven dried at 60 °C for 48 hours before samples were analysed in
137 the laboratory. For this analysis, a piece of the stem was scraped to remove any
138 external dirt, and a small sliver of the clean stem, approximately 0.60 mg in weight,

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taken with a scalpel and put into a tin capsule. The tin capsule was then combusted in a Flash EA[®] 1112 series elemental analyzer (Thermo Finnigan, Italy). The CO₂ produced was fed into a Delta[®] Plus XP IRMS isotope ratio mass spectrometer (Thermo Electron, Germany) via a Conflo III[®] gas control unit (Thermo Finnigan, Germany). The carbon isotope ratio (¹³C/¹²C) output was then reported relative to the PDB standard. The resulting delta notation, δ¹³C, was determined using the following equation (Kloppel *et al.* 1998):

$$\delta^{13}C(\text{‰}) = \frac{R_{\text{sample}} - R_{\text{standard}}}{R_{\text{standard}}} \times 1000$$

Hydrological monitoring

To enable an accurate understanding of soil water regime within the plots, a hydrologic model based on Gowing & Youngs (1997) was used. The model was built from inputs of water-table-depth behaviour in the field, topographic variation, soil characteristics and meteorological data.

The water-table depth was monitored from an array of dip wells, supported by automatic logging ‘divers’ (Eijkelkamp[®], The Netherlands). The dip wells were read manually every two weeks, while automatic divers in a sub-sample of wells were set to read every four hours for 15 months’ duration. We confirmed the suitability of water table depth as a proxy for assessing water availability in our sites, by regressing it against volumetric water content during the monitoring season.

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4 163 To account for micro-topographical differences, topography was surveyed at
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6 164 all quadrat and dip well locations using a Leica Geosystems® TPS300 (Switzerland)
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8 165 total station device.
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12 167 Using the hydrological monitoring from dipwells and divers, it was possible to
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14 168 interpolate the water-table depths for each quadrat location. The mean water-table
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16 169 depths of each quadrat from the spring season, was then weighted by the % presence
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18 170 of the respective species present, to get its hydrological metric (i.e. mean water-table
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20 171 depth).
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26 27 173 **Phylogeny** 28 29 174

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31 175 The Restionaceae comprise 350 species, 342 of which form a clade that is
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33 176 endemic to the Cape Floristic Region. The phylogeny of Restionaceae used in this
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35 177 study was built on the basis of *rbcl* markers and morphological traits (Hardy *et al.*
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37 178 2008) and includes 94% of the Cape clade, resolved to species level.
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42 43 180 **Statistical analysis** 44 45 181

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47 182 Initially, the botanical data collected for each site were tested for spatial
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49 183 community structure using Stone and Roberts' C-score in the EcoSim software
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51 184 (Gotelli & Entsminger 1997). The C-score measures the average number of
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53 185 "checkerboard units" (i.e. species pairings that do not occur together in a quadrat)
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55 186 between all possible pairs of species. In a competitively structured community, the C-
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3 187 score should be significantly larger than expected by chance. 10,000 simulations with
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5 188 a random seed of 10 were run and compared against the actual field distribution.
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10 190 The species-level $\delta^{13}\text{C}$ values were analysed using one-way and when
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12 191 appropriate two-way analysis of variance. The difference in $\delta^{13}\text{C}$ values between male
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14 192 and female plants was tested using a two-sample t-test. All the analysis was conducted
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16 193 using Statistica[®] Release 8.0 software.
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22 195 The correlation between mean water table depth and $\delta^{13}\text{C}$ was measured by
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24 196 the method of phylogenetically independent contrasts, an approach often used for
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26 197 cross-species comparisons (Felsenstein 1985). Such analysis needs to be conducted to
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28 198 account for the potential similarity of closely related species (phylogenetic history);
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30 199 which otherwise prevents individual species from being statistically independent data
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32 200 points (Felsenstein, 1985). The comparative contrast was implemented using the
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34 201 CACTUS 1.13 (Comparative Analysis of Continuous Traits Using Statistics; Schwikl
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36 202 & Ackerly 2001) software.
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43 204 **RESULTS**
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48 206 **Species**
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53 208 There was an average of 2.1 Restionaceae species per quadrat at each site. The
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55 209 botanical survey at Riverlands found 18 Restionaceae species, of which we studied
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57 210 the ten most prevalent species for logistical reasons. Of these ten species it was only
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59 211 possible to identify gender in individuals of five species. At New Years Peak a total of
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12 species were found, of which we used 11 in our study (Table 1). Gender could be scored in eight of the species.

<Table 1>

Community structure

Statistical analysis of the field distribution of Restionaceae using Robert's C-score for co-occurrence (Gotelli & Entsminger 1997) revealed significant structuring, compared to what would be expected by chance (with observed values > simulated values at $p < 0.001$).

Water regime

The water-table depths in individual quadrats, as modelled from biweekly monitoring, ranged from -0.20 to 0.65 m for New Years Peak site and between 0.30 and 1.46m for Riverlands. The mean spring water-table depths for the quadrats occupied by the studied Restionaceae species are given in Table 1.

$\delta^{13}\text{C}$ values

$\delta^{13}\text{C}$ values observed ranged from -26.9 ‰ (*Calopsis viminea*) to -28.77 ‰ (*Willdenowia arescens*) in Riverlands and from -25.2 ‰ (*Anthochortus crinalis*) to -27.2 ‰ (*Elegia coleura*) in New Years Peak. The standard errors within each species ranged from 0.20 - 0.51 (Riverlands) and 0.22 – 0.62 (New Years Peak). One way

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3 235 analysis of variance showed there were significant differences in $\delta^{13}\text{C}$ signatures
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6 236 among the species at both sites (Table 2).
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15 240 $\delta^{13}\text{C}$ values for females were generally more negative, -27.5 ‰ (New Years
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17 241 Peak) and -28.2 ‰ (Riverlands) than for their male counterparts -26.7 ‰ (New Years
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19 242 Peak) and -27.9 ‰ (Riverlands). Combined analysis of the difference between males
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21 243 and females of all species at the two sites, showed this difference was significant (two
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23 244 sample t-test $p < 0.001$)
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29 246 Two-way analysis of variance, for species whose gender was known, also
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31 247 showed significant differences between the species (both sites) as well as genders of
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33 248 the plants (only at New Years Peak). However, there was no significant interaction
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35 249 between species and gender at any of the sites (Table 3).
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46 253 **$\delta^{13}\text{C}$ versus water table depth**
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51 255 Taking account of phylogenetic relationships, through PICS analysis, mean
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53 256 water-table depth was plotted against $\delta^{13}\text{C}$ (Figure 1). The results show that there was
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55 257 no significant linear correlation between species' $\delta^{13}\text{C}$ and their mean water-table
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57 258 depth along the hydrologic gradient at either site ($r^2 = 0.23$ $p = 0.34$, New Years Peak
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59 259 and $r^2 = 0.06$ $p = 0.60$, Riverlands).
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260

261 <Figure 1>

262

263 **DISCUSSION**

264

265 Soil moisture availability is a key environmental variable affecting plant
 266 distribution and coexistence along hydrological gradients (Silvertown 2004;
 267 Engelbrecht *et al.* 2007). In this study, we found that Restionaceae species showed
 268 niche segregation, as quite unrelated species also do for example in wet meadows in
 269 England (Silvertown *et al.*, 1999). Intrinsic water use efficiency is expected to vary
 270 with hydrological conditions and has implications for plant distribution and survival
 271 strategies. For example, where water supply is abundant, plants with a non-
 272 conservative water-use strategy seem to be the most successful (Chen *et al.* 2007). In
 273 the past, $\delta^{13}\text{C}$ has been used as a surrogate for WUE and so we tested whether there
 274 was a correlation between $\delta^{13}\text{C}$ and hydrological conditions measured in the field.

275

276 The Restionaceae $\delta^{13}\text{C}$ values observed in these sites, shown in Table 1 (from
 277 -25.2 ‰ to -28.8 ‰) are well within the range of what would be expected of C_3 plants
 278 (Griffiths *et al.* 2000; Linder 2005). The larger values of $\delta^{13}\text{C}$ at New Years Peak than
 279 at Riverlands (Table 1) could be attributed to differences in altitude, precipitation or
 280 temperature (Panek & Waring 1995; Van de Water 2002; Zheng & Shangguan 2007),
 281 since New Years Peak is 965 m higher than Riverlands and has rainfall of 1500 mm
 282 per year, four times that of Riverlands.

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284 The more negative $\delta^{13}\text{C}$ signature that we found in female plants is in
285 agreement with previous findings in other species (Dawson & Ehleringer 1993;
286 Dudley 2006). The gender difference is usually interpreted as evidence of water use
287 efficiency, which is improved through reduction of both water loss and carbon gain
288 (Dawson and Geber 1999). In this context, females of dioecious species often show
289 higher reproductive effort than males because females allocate more biomass to
290 reproduction in the form of flowers and fruits (Correia & Diaz Barradas 2000) thereby
291 reducing their WUE.

292

293 Stomatal limitation of photosynthesis mediated by water stress is currently
294 accepted as one of the main limitations to plant productivity in dry-land ecosystems
295 (Diaz-Espejo *et al.* 2007). The close relationship between WUE and transpiration
296 efficiency, is also known to be reflected in carbon isotope discrimination (Farquhar *et*
297 *al.* 1982). Thus, the significant differences in the $\delta^{13}\text{C}$ values observed (Tables 2 and
298 3) amongst species within a single family and having similar life form, suggests that
299 they have different stomatal behaviour. Physiologically, this is a strong suggestion of
300 environmental control (in this case by water) on plants' response (stomatal behaviour
301 and hence carbon isotope discrimination). Ecologically, the implication goes even
302 further: it could be regulating species distribution pattern by modifying their
303 competitive interaction. For example, Mole *et al.* (1994) found that prairie grass
304 *Agropyron smithii*, which has a more negative mean $\delta^{13}\text{C}$, showed further range
305 expansion than other coexisting species following drought. Even the level of diversity
306 in leaf gas-exchange regulation has been associated with high species diversity in
307 tropical rainforest (Bonal & Guehl 2001).

308

309 $\delta^{13}\text{C}$ values and water-table depth

310

311 Previous studies in the investigation of $\delta^{13}\text{C}$ and water availability, have been
 312 conducted either at coarse continental, altitudinal or large landscape vegetation units
 313 (e.g. Stewart *et al.* 1995; Kloeppel *et al.* 1998; Van de Water *et al.* 2002; Wang *et al.*
 314 2005; Chen *et al.* 2007) or only between genotypes of a species level (e.g. Yoneyama
 315 & Ohtani 1983; Farquhar & Richards 1984; Martin & Thorstenson 1988). Only very
 316 few have looked at subtle topo-edaphic gradients within a landscape scale (Peñuelas
 317 *et al.* 1999; Bai *et al.* 2008). Furthermore, the most prevalent generalization from the
 318 above studies has been for plant $\delta^{13}\text{C}$ values to be positively correlated with water
 319 availability, which was often expressed as precipitation. In these studies, more
 320 negative values of $\delta^{13}\text{C}$ have been interpreted as indicators of greater enzymatic
 321 discrimination, implying higher stomatal conductance and hence lower water use
 322 efficiency (e.g. Farquhar & Richards 1984).

323

324 In woody fynbos plants, species in the same genera show similar drought
 325 vulnerability curves, regardless of whether or not they were wetland adapted or
 326 dryland adapted species. This suggests strong phylogenetic constraint on xylem
 327 function (Aston 2007). In this study even though we accounted for phylogeny of our
 328 species, we still found no significant correlation between the $\delta^{13}\text{C}$ and the subtle
 329 hydrologic gradient, even as the species segregated along the gradient (Figure 1).

330

331 Although this comes as a surprise, it reinforces recent views by Seibt *et al.*
 332 (2008), on the insufficiency of $\delta^{13}\text{C}$ values alone as reliable surrogates to WUE to
 333 explain plant response distributions in relation to water regime, particularly in natural

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3 334 field populations. This is because a number of other environmental interactions could
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5 335 influence $\delta^{13}\text{C}$ (Vitousek *et al.* 1990; Griffiths *et al.* 2000). For example, Bai *et al.*
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8 336 (2008) found the relationship between $\delta^{13}\text{C}$ and water regime was the opposite of
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10 337 what would be expected and they suggested that another environmental factor i.e.
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12 338 nitrogen availability could be a mediating factor between water availability and plant
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14 339 $\delta^{13}\text{C}$ response. In our case, we have accounted for the factors: species, gender,
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16 340 phylogeny, and water regime, but not the seasonality component i.e. phenology of the
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18 341 species. It is known that plant carbon isotope discrimination varies with the plant's
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20 342 stage of development and season (e.g. Mole *et al.* 1994; Zhao *et al.* 2004; Chen &
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22 343 Chen 2007). Smedley *et al.* (1991) showed there is a general trend for a decrease of
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24 344 carbon isotope discrimination as the growing season progresses i.e. a response to
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26 345 increase in evaporative demand and decrease in soil moisture. Hence early flowering
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28 346 species discriminate more than later flowering ones, because such species are active
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30 347 during the initial, less water-stressed months of the growing season and tend to use
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32 348 water less efficiently. As such phenology is a possible explanation for the differences
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34 349 we saw in our species, particularly as Restionaceae flower year-round, with different
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36 350 species flowering in different months (Linder 2002).
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46 352 **ACKNOWLEDGEMENTS**
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547 **Table 1.** Species studied at each site, % presence, height, $\delta^{13}\text{C}$ and mean water table
548 depth

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550 **Table 2.** Analysis of variance results for $\delta^{13}\text{C}$ among species of Restionaceae at New
551 Years Peak and Riverlands

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553 **Table 3.** Two-way analysis of variance of $\delta^{13}\text{C}$ for Restionaceae by species and
554 gender at New Years Peak and Riverlands

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556 **Figure 1.** Phylogenetically independent contrasts of mean water table depth and $\delta^{13}\text{C}$
557 for species sampled at New Years Peak and Riverlands.

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Table 1. Species studied at each site, % presence, height, $\delta^{13}\text{C}$ and mean water table depth

Site	Species	% Presence	Average height (m)	Mean $\delta^{13}\text{C}$ (‰)	Mean Water table (m)
Riverlands	<i>Calopsis viminea</i>	10.2	0.48	-26.9	0.54
	⁺ <i>Cannomois acuminata</i> [‡]	12.1	0.78	-27.5	0.86
	<i>Chondropetalum nudum</i>	21.0	0.74	-28.7	0.63
	<i>Elegia filacea</i>	28.2	0.68	-28.0	0.43
	<i>Hypodiscus willdenowia</i> [‡]	13.1	0.45	-27.8	0.57
	<i>Ischyrolepis capense</i>	10.5	0.50	-28.2	0.58
	⁺ <i>Ischyrolepis monanthos</i>	12.5	0.60	-27.5	0.83
	⁺ <i>Staberoha distachyos</i>	24.9	0.65	-28.5	0.68
	⁺ <i>Thamnochortus punctatus</i>	38.4	0.91	-28.2	0.83
	⁺ <i>Willdenowia arescens</i>	17.0	0.68	-28.8	0.78
New Years Peak	<i>Anthochortus crinalis</i>	35.7	0.33	-25.2	0.07
	⁺ <i>Elegia coleura</i>	13.2	0.48	-27.2	0.15
	<i>Elegia filacea</i>	23.8	0.27	-26.3	0.36
	<i>Elegia neesii</i>	50.6	0.49	-26.5	0.25
	⁺ <i>Ischyrolepis curviramis</i> [§]	26.8	0.20	-26.8	0.34
	⁺ <i>Restio pedicellatus</i>	10.6	0.44	-27.1	0.33
	⁺ <i>Restio bolusii</i>	23.8	0.31	-25.9	0.30
	⁺ <i>Hypodiscus arescens</i>	<10.0	0.51	-28.1	na
	⁺ <i>Restio obscurus</i>	<10.0	0.55	-27.5	na
	⁺ <i>Staberoha cernua</i>	<10.0	0.24	-26.3	na
	⁺ <i>Ceratocarium fimbriatum</i> [§]	<10.0	0.69	-27.7	na

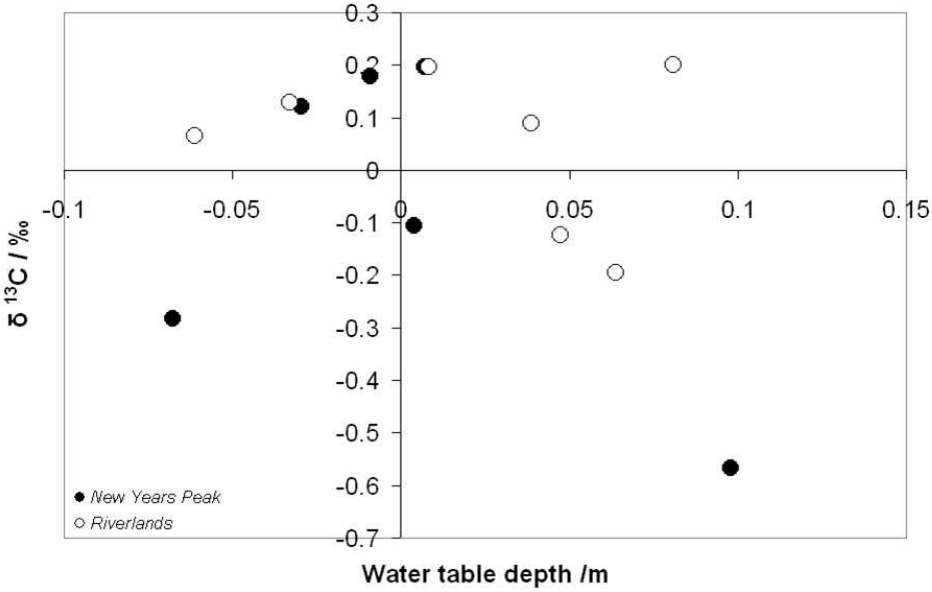
⁺ species in which males and females were distinguished; [‡]species missing from phylogeny [§] there were only 5 replicates for the species (10 for the other species); ^{na} no sufficient hydrologic data available due to <10 % presence.

Table 2. Analysis of variance results for $\delta^{13}\text{C}$ among species of Restionaceae at New Years Peak and Riverlands

Site	Source	df	MS	F	p
New Years Peak	Species	10	6.80	5.71	<0.001
	Error	89	1.19		
Riverlands	Species	9	3.54	3.58	<0.001
	Error	90	0.99		

Table 3. Two-way analysis of variance of $\delta^{13}\text{C}$ for Restionaceae by species and gender at New Years Peak and Riverlands

Site	Source	df	MS	F	p
New Years Peak	Species	7	3.94	4.13	<0.002
	Gender	1	4.52	4.73	0.04
	Species x Gender	7	0.98	1.03	0.43
	Error	37	0.96		
Riverlands	Species	4	3.40	2.98	0.03
	Gender	1	1.42	1.24	0.27
	Species x Gender	4	2.78	2.43	0.06
	Error	40	1.14		



Phylogenetically independent contrasts of mean water table depth and δ¹³C for species sampled at New Years Peak and Riverlands.
254x190mm (96 x 96 DPI)